


RESEARCH ARTICLE

Soil flooding filters evolutionary lineages of tree communities in Amazonian riparian forests

Sthefanie do Nascimento Gomes de Souza^{1,2}  | Darlisson Mesquita Batista³ | Adriano Costa Quaresma^{2,4} | Ana Luiza Costa² | Layon Oreste Demarchi² | Bianca Weiss Albuquerque² | Viviane Pagnussat Klein² | Gildo Feitoza² | Angélica Faria de Resende^{2,5} | Gisele Biem Mori² | Florian Wittmann⁴ | Leidiane Leão Oliveira⁶ | Amanda Frederico Mortati⁷ | Alan Cavalcanti da Cunha⁸ | Jochen Schongart² | Aline Lopes^{2,9} | Maria Teresa Fernandez Piedade² | Thiago André¹⁰

¹Postgraduate Program in Ecology, National Institute for Amazon Research (INPA), Manaus, Brazil

²Ecology, Monitoring and Sustainable Use of Wetlands (MAUA Research Group), National Institute for Amazon Research (INPA), Manaus, Brazil

³Postgraduate Program in Biodiversity, Federal University of Western Pará, Santarém, Brazil

⁴Institute of Technology (KIT), Karlsruhe, Germany

⁵Forest Sciences Department, ESALQ/USP, University of São Paulo, São Paulo, Brazil

⁶Institute of Water Sciences and Technology, Federal University of Western Pará, Santarém, Brazil

⁷Center for Sustainable Development, University of Brasília, Brasília, Brazil

⁸Program in Environmental Science (PPGCA), Federal, University of Amapá (UNIFAP), Macapá, Brazil

⁹Researcher at the Cesumar Institute of Science, Technology and Innovation (ICETI), Maringá, Brazil

¹⁰Department of Botany, Institute of Biological Sciences, University of Brasília, Brasília, Brazil

Correspondence

Sthefanie do Nascimento Gomes de Souza, Postgraduate Program in Ecology, National Institute for Amazon Research (INPA), Av. André Araújo, 2936-Petrópolis, Manaus 69067-375, AM, Brazil.
Email: sthefaniegsouza@gmail.com

Funding information

Conselho Nacional de Desenvolvimento Científico e Tecnológico (CNPq), Grant/Award Number: 441462/20170, 441481/20175, 441498/20175 and 441520/20170; Universidade de Brasília (UnB), Grant/Award Number: (DPI/DPG/BCE01/2024); Fundação de Amparo Pesquisa do Estado de São Paulo- FAPESP, Grant/Award Number: 19/240495 and 22/146050

Abstract

Inundations in Amazonian black-water river floodplain result in the selection of different tree lineages, thus promoting coexistence between species. We investigated whether Amazonian tree communities are phylogenetically structured and distributed along a flooding gradient from irregularly flooded forests along streams embedded within upland (*terra-firme*) forest to seasonally flooded floodplains of large rivers (*igapós*). Floristic inventories and hydrological monitoring were performed along the Falsino River, a black-water river in the eastern Amazon within the Amapá National Forest. We constructed a presence-and-absence matrix and generated a phylogeny using the vascular plant database available in GenBank. We calculated the standardized values of the metrics of phylogenetic diversity (ses.PD), average phylogenetic distance (ses.MPD), and average nearest-neighbor distance (ses.MNTD) to test whether the history of relationships between species in the community is influenced by inundation. We used the phylogenetic endemism (PE) metric to verify the existence of taxa with restricted distribution. Linear regressions were used to test whether phylogenetic metrics have a significant relationship with the variables: maximum flood

This is an open access article under the terms of the [Creative Commons Attribution](https://creativecommons.org/licenses/by/4.0/) License, which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

© 2024 The Author(s). *Ecology and Evolution* published by John Wiley & Sons Ltd.

height, maximum water table depth, and maximum flood amplitude. The results show that forests subject to prolonged seasonal flooding have reduced taxon richness, low phylogenetic diversity, and random distribution of lineages within communities. On the other hand, *terra-firme* riparian forests showed higher rates of taxon richness, diversity, and phylogenetic dispersion, in addition to greater phylogenetic endemism. These results indicate that seasonal and predictable soil flooding filters tree lineages along the hydrographic gradient. Different adaptations to root waterlogging are likely requirements for colonization in these environments and may represent an important factor in the diversification of tree lineages in the Amazon biome.

KEYWORDS

endemism, overdispersion, phylogenetic diversity, randomness, wetlands

TAXONOMY CLASSIFICATION

Biodiversity ecology, Botany, Community ecology, Conservation ecology, Conservation genetics, Ecosystem ecology, Evolutionary ecology, Phylogenetics

1 | INTRODUCTION

The Amazon basin extends over 6.8 million km² and presents a high richness and diversity of species (Antonelli & Sanmartín, 2011; Junk et al., 2011; Myers et al., 2000; ter Steege et al., 2013). It is a vast and interconnected system composed of different landscapes of upland (i.e., *terra-firme*) forests and wetlands, linked by an extensive drainage system that has undergone significant changes throughout its history (Hoorn et al., 2010). This basin regulates a wide range of aquatic and terrestrial environments that are characterized by a diverse array of tree species adapted to different soil conditions and hydrological gradients (Junk et al., 1989, 2011; Prance, 1980; Salati & Vose, 1984). Here, we divide Amazonian forests into *terra-firme* (upland) forests, which lie above the maximum flood level of rivers, and lowland floodplains that are either seasonally or permanently inundated by its marginal water body (Bredin et al., 2020; Sioli, 1985; Sombroek, 2000). The oligotrophic and sediment-poor transparent or black water drain the oldest geological formations of the Amazon, predominantly formed during the Tertiary period, with alluvial sediments originating from Precambrian-aged rocks due to the erosion of the Guiana and Central Brazilian Shields (Sioli, 1985; Irion et al., 1997). These types of flooded forests (*igapós*) exhibit nutrient scarcity and intense weathering, resulting in high endemism and low species richness compared to the *várzeas*, floodplain forests inundated by the more recent white-water sediment-rich rivers (Ayres, 1986; Fittkau, 1971; Junk et al., 1989; Kubitzki, 1987; Prance, 1979). Many tree species in *igapó* riparian forests are ecologically specialized to this ecosystem, leading to distinct floristic compositions when compared to *várzea* and *terra-firme* riparian forests (Householder et al., 2024; Kubitzki, 1989), with a unique evolutionary signature of its phylogenetic composition (Luize et al., 2024).

Riparian zones form at the dynamic contacts between terrestrial and aquatic ecosystems, marginal to water bodies, showcasing diverse patterns in relation to the organization, diversity, and dynamics

of amphibious communities (Naiman & Décamps, 1997). Riparian vegetation varies with soil water saturation, which increases from low-order rivers in the catchment area toward high-order rivers in the lowlands with simultaneously changing flood-pulse patterns (Gregory et al., 1991; Junk et al., 1989, 2011). The flood pulse modifies the environment quickly in time and space (Junk et al., 1989). Seasonal flooding and deeper water tables in lowland regions, where flood pulses are regular and of greater amplitude (i.e., *igapó*), result in the predominance of species adapted to the long-term flooding of several months (Koga et al., 2024; Parolin, 2009; Parolin & Wittmann, 2010; Piedade et al., 2013; Wittmann et al., 2013). In contrast, forests along streams and small rivers embedded within uplands (*terra-firme* riparian forests), where the water table is shallow and closely related to surface water levels, experience stochastic, unpredictable, and low-amplitude flood regimes. These environments harbor a higher diversity of species adapted to heterogeneous soil composition, facing different selective pressures from those found in large-river floodplains. This results in a distinct species composition along the riparian zone of freshwater bodies (Junk et al., 2011; Kubitzki, 1989; Prance, 1980). Upland riparian environments generally feature poorly drained soils compared to those of large-river floodplains (Drucker et al., 2008; Schietti et al., 2014), requiring adaptations from colonizing trees. The various types of tree species adaptations to the floodable areas of the Amazon basin are well documented (e.g., Da Silva et al., 2023; Ferreira et al., 2010; Junk et al., 2015; Parolin, 2009; Piedade et al., 2013).

Riparian forests along low-order rivers bridge the gap between the *terra-firme* and the alluvial plains of large rivers, with many tree species found in both floodable and upland areas. The concept of floodplain species colonization, proposed by Wittmann et al. (2010), suggests that the colonization of tree species in the alluvial plains of large rivers originated from the species pool of the *terra-firme* over evolutionary periods, as the species gradually developed increasingly sophisticated adaptations to tolerate longer floods. However,

further phylogenetic studies along the river continuum as suggested by Vannote et al. (1980) are necessary to provide additional support for the idea of species colonization in floodplains, highlighting the evolutionary processes shaping the diversity and distribution of Amazonian tree species across different habitats (Wittmann et al., 2022).

The complex interaction between hydrological factors that result in the flooding gradient plays a determining role in the organization of the different phytophysiognomies in the Amazon, considering the remarkable ability of plants to adapt to changing environmental conditions (Junk & Piedade, 1997; Parolin, 2009). Tree species of various taxonomic groups may exhibit convergent morphological and physiological adaptations, such as adventitious and aerenchyma roots, stem hypertrophy, lenticels, and reduced photosynthesis, in response to water stress (Ferreira et al., 2006; Koga et al., 2024; Kozłowski, 1997; Lobo & Joly, 1995; Wittmann & Parolin, 1999); whereas species composition and richness vary significantly both among and within floodable forest communities in the Amazon (Ferreira et al., 2005; Hoorn et al., 2017; Lobo et al., 2019; Rocha et al., 2019).

Despite its vastness and species richness, knowledge gaps exist regarding areas with high phylogenetic diversity (PD) (Webb, 2000) and phylogenetic endemism (PE) in Amazonian riparian zones (Rosauer et al., 2009). Phylogenetic diversity quantifies the evolutionary history of organisms based on ancestral relationships and is considered crucial for identifying areas with high diversity, while species endemism indicates the presence of taxa with restricted spatial distribution (Faith, 1992; Magurran, 2004; Owen et al., 2019; Ricotta, 2005; Vane-Wright et al., 1991; Webb, 2000). Areas with high phylogenetic endemism play a vital role in biodiversity conservation (Ferrier et al., 2000; Pressey et al., 1994) and act as nature reserves for maximizing local phylogenetic diversity (Faith et al., 2004; Forest et al., 2007). These areas are extremely important in the preservation of species and in ensuring the maintenance of evolutionary diversity, considering the alarming loss of biodiversity at the global level (Cadotte et al., 2008; Dirzo & Raven, 2003; Isbell et al., 2023).

Factors, such as habitat characteristics and species traits, can influence patterns of phylogenetic richness and diversity (Cadotte et al., 2010; Ricklefs et al., 2014; Stropp et al., 2011). Environmental factors can function as filters, selecting species that are phylogenetically close or distant, and which may share similar functional characteristics to exploit the same resources in different ways (Fukami et al., 2005; Pausas & Verdú, 2010; Webb, 2000). This means that the coexistence of species can involve both those that are evolutionarily close (clustered) and those that are distant (overdispersed), or it can result in randomly distributed communities in space. Understanding these phylogenetic patterns is critical in order to obtain more accurate estimates of how plant communities are structured in space and time (Fine & Kembel, 2011; Gerhold et al., 2018; Kubitzki, 1987; Muscarella et al., 2020; Wittmann et al., 2010).

Recently, Luize et al. (2024) demonstrated that numerous tree lineages serve as significant indicators of geographic regions in

Amazonia. The authors emphasized that *terra-firme* and white-water floodplain forests (*várzea*) exhibit the most variable phylogenetic composition, while white-sand forests (*campinarana*) and black-water floodplain forests (*igapó*) tend to have less variable phylogenetic composition. Here, we focus on a local river basin scale, where we evaluated two hypothetical scenarios to unravel historical evolutionary processes: (1) a scenario where only a few tree lineages have evolved traits that enable survival in floodable areas, resulting in a phylogenetically clustered community; or (2) a scenario where independent lineages of tree species have repeatedly colonized these environments, leading to convergences of evolutionary strategies for surviving flooding and thus resulting in reduced phylogenetic clustering along a gradient from upland small-streams riparian forests (*terra-firme*) to large-river floodplain forests (*igapós*). Therefore, the main objective of this study was to analyze the diversity and phylogenetic structure of tree communities, ranging from riparian forests adjacent to small streams on *terra-firme*, i.e., upper lands, with brief, shallow, and unpredictable flooding, to large rivers on *igapó*, i.e., lower lands, with prolonged, deep, and predictable flooding. The analysis aimed to establish relationships between these environments and the hydrological gradients of flooding, above and below the soil surface, in order to enhance our understanding of the organization of tree communities in Amazonia, which thrive amidst a mosaic of these distinct conditions.

2 | MATERIALS AND METHODS

2.1 | Study site

This study was conducted in the oligotrophic black-water sub-basin of the Falsino River (0°55' N and 51°35' W), situated adjacent to the Amapá National Forest (FLONA Amapá), in the state of Amapá, Brazil (Figure 1). Covering an area of 460,353 ha, FLONA Amapá borders the Amapá State Forest to the east and the Montanhas do Tumucumaque National Park to the west, collectively forming the Amapá biodiversity corridor, which spans over 70% of the state's total area (ICMBIO, 2014). The perimeter of FLONA Amapá is delineated by the Araguari River to the south, the Falsino River to the east, and the Mutum River to the west. The Falsino River, a major tributary of the Araguari River, flows more than 70 km upstream from the Porto Grande hydroelectric dam, remaining in its natural state and accessible only with authorization for scientific research or tourism (ICMBIO, 2016). With its most distant headwaters in the Guiana Shield, this black-water river carries a low sediment and nutrient load (Junk et al., 2015). Historically, the Araguari River discharged into the Atlantic Ocean; however, due to erosion of the Urucurituba channel, most of its water now diverts to the Amazon River (Da Silva et al., 2023; dos Santos et al., 2018).

The climate of the FLONA Amapá region is tropical hot-humid type, with an average temperature of 27.6°C in Macapá and an average annual rainfall of 2,284 mm in the Serra do Navio region,

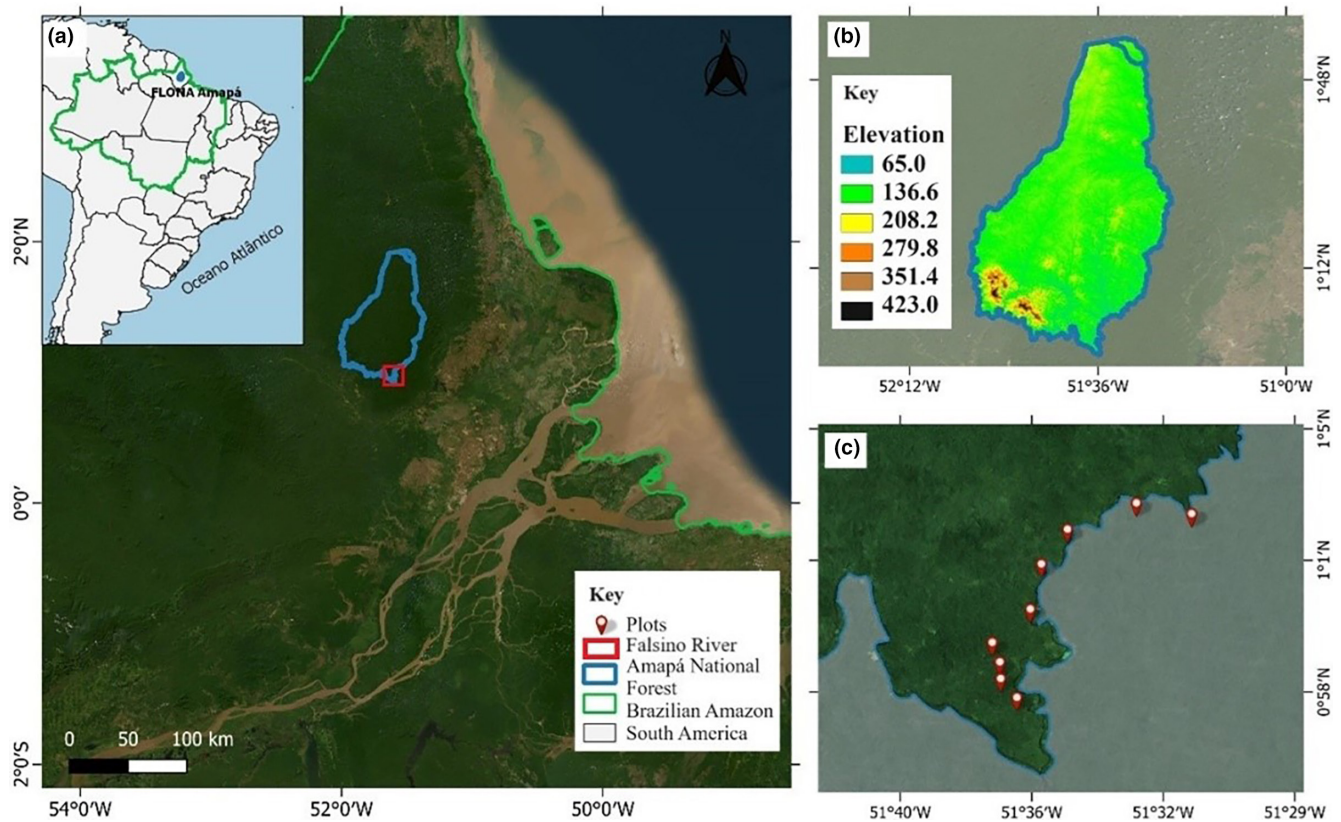


FIGURE 1 (a) Location of the study area in the Amapá National Forest; (b) Digital surface elevation model (MDE Copernicus); and (c) Location of the plots. Source: Geoprocessing and Modeling Center – MAUA /INPA Group.

the closest municipalities with available data (SUDAM, 1984). Situated in the Guiana Shield region, FLONA Amapá features geological formation dating from the Precambrian (>1,000 Mya) and Paleoproterozoic periods (2,300–2,050 Mya) (Gómez et al., 2019). The terrain ranges from 50 m above sea level (a.s.l.) to peaks reaching 460 m a.s.l. in the southern part of the protected area (ICMBIO, 2016). The vegetation is heterogeneous, encompassing *terra-firme* forests, lowlands, flooded forests, and transitional zones between forests and savannas (ICMBIO, 2014).

2.2 | Sampling design

The floristic inventory included trees with a diameter at breast height (DBH) of ≥ 10 cm and was conducted in August 2018. Species identification in the field was carried out with the assistance of a botany taxonomist. Additionally, we collected botanical vouchers in the field and deposited them at the INPA (National Institute for Amazon Research) herbarium in Manaus-AM-Brazil for final identification. The sub-basin survey comprised nine plots of 0.5 ha (100 m \times 50 m), each of them subdivided into eight sub-plots of 625 m² to minimize within-plot variation in altitude, soil composition, and water table depth. The plots were distributed along the sub-basin's different flood regimes as follows: three in first- or second-order small streams in the upper stretches

(*terra-firme* riparian forests), three along third-order rivers in the middle stretches (intermediate riparian forests), and three in the floodplains of the main fourth-order river in the lower stretches (*igapó* riparian forest; Figure 2). Plot locations were selected based on topography, maintaining a spacing of over 1 km between them to maximize sampling independence. The standardization of the sampling units enabled capturing environmental variation along the fluvial continuum from the headwaters to the mouth of the Falsino River.

2.3 | Hydrological gradient monitoring

To measure flood height above the soil surface, limimetric rulers were installed along the channel bank closest to the plot, affixed to healthy, resistant trees using nails. These aluminum rulers, 3 m in length, were graduated every 1 cm and numbered every 2 cm. Monthly water level monitoring was conducted over a 1-year period (Belúcio, 2020). In cases where flooding exceeded 3 m, a spare ruler was used alongside the installed one for measurement. Water table depth was determined by collecting water fluctuation data from piezometric wells with depths of 3–6 m, present in each plot. Flood amplitude was estimated based on the difference between maximum flood height and maximum water table depth using hydrological data monitoring.

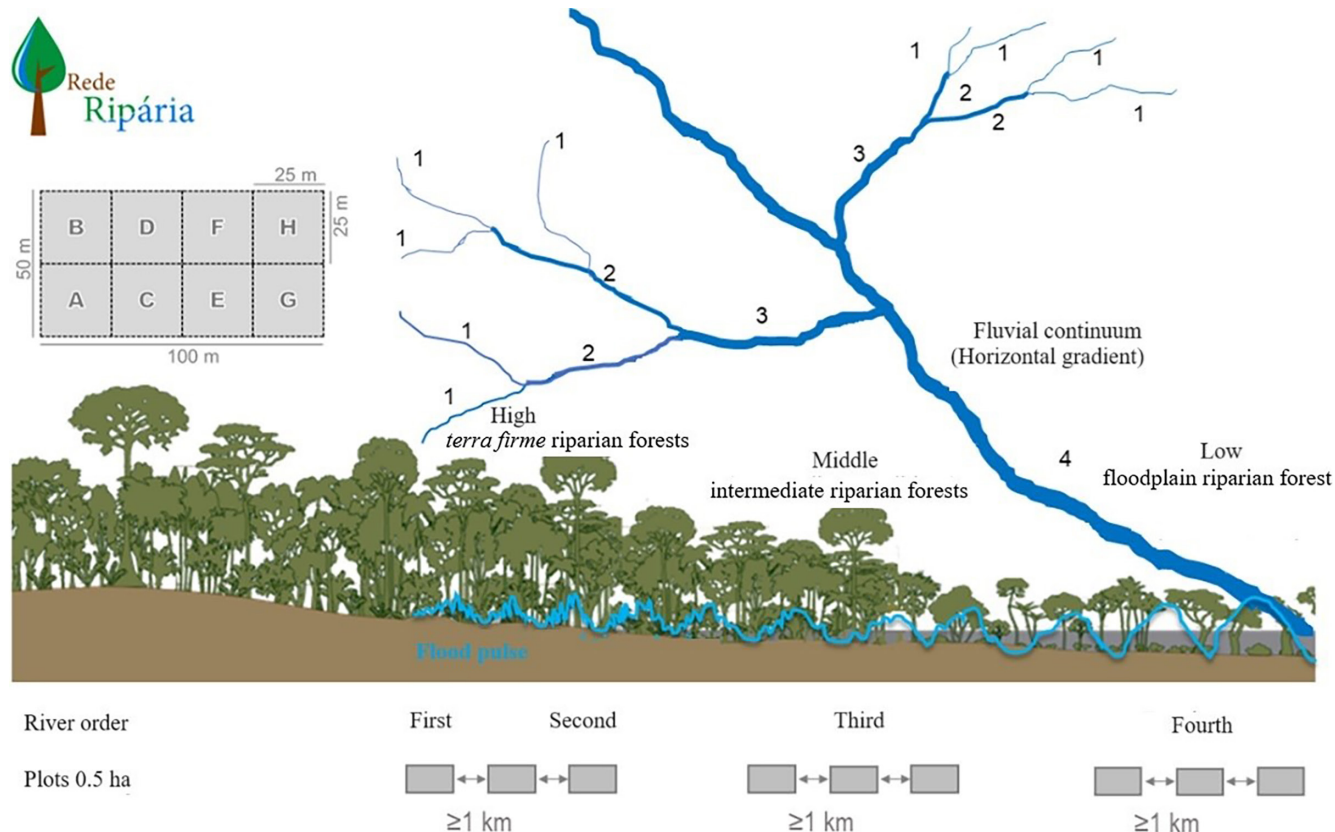


FIGURE 2 Sampling design of the distribution of plots of 0.5 ha (100 m x 50 m) subdivided into eight subplots of 625 m² allocated along the river orders; The first- to second-order rivers (shallows) are represented by riparian forests in the upper section, the third-order rivers have intermediate conditions represented by the middle section, and the fourth-order rivers have flooded forest formations (*igapós*) represented by the low section. The river continuum corresponds to all sampling carried out from the headwaters to the mouth of the sub-basin. Source: Jochen Schöngart/Rede Ripária.

2.4 | Species relationships and phylogenetic community structure

The V.PhyloMaker package (Jin & Qian, 2019) was employed to reconstruct the phylogenetic hypothesis of tree species based on the presence-and-absence matrix of species inventoried. V. PhyloMaker utilizes an updated and expanded version of the GBOTB (GenBank taxa with a backbone provided by Only Tree of Life version 9.1) mega-phylogeny developed by Smith and Brown (2018). This mega-phylogeny encompasses all existing seed plant families (74,533 species) according to the APG IV system (Angiosperm Phylogeny Group 2016), rendering it the most extensive dated phylogeny for vascular plants available (Jin & Qian, 2019). Species absent in the GBOTB mega-phylogeny but present in our dataset were added to their respective genera using scenario 3 with the scenarios = "S3" function implemented in V.PhyloMaker. The resulting phylogeny includes 33.3% of unresolved taxa (polytomies).

Species richness was calculated as the number of species in each plot, and raw values of phylogenetic diversity (PD) (Faith, 1992), mean phylogenetic distance (MPD), and mean nearest taxon distance (MNTD) were determined. To address a possible bias related to these three phylogenetic metrics and raw values of PD, MPD, and MNTD is their sensitivity to species richness present in the community

(Cadotte et al., 2010; Sandel, 2018; Tucker & Cadotte, 2013). We removed the richness effect from the raw metric values (PD, MPD, and MNTD) (Kembel et al., 2010; Miller et al., 2017; Webb et al., 2002) and used the null model of the standardized effect size (ses.PD, ses.MPD, and ses.MNTD). The phylogenetic structure analyses were conducted using the 'picante' package, version 1.8.2 (Kembel et al., 2010). The null model of standardized effect size (ses.PD, ses.MPD, and ses.MNTD) was applied. The null model involved performing 999 permutations using the "independent swap" method, implemented in the R 'picante' package, following manual instructions. This method randomly rearranges the labels of the taxa in the phylogenetic distance matrix and allows for the calculation of the mean and standard deviation (SD) of the values obtained from the models generated by these permutations (Kembel et al., 2010). The "independent swap" model assumes that the ability of a species to colonize a plot is proportional to its frequency in the database, with each species having an equal probability of occurring in each plot (Gotelli & Graves, 1996; Hardy, 2008). Positive values of ses.PD suggest greater diversity, ses.MPD and ses.MNTD with $p \geq .95$ indicate phylogenetic overdispersion, negative values indicate lower diversity, $p \leq .05$ indicate phylogenetic clustering, and values around zero and without significance (p between .05 and .95) indicate a random phylogenetic pattern.

In addition to standardized ses.PD, ses.MPD, and ses.MNTD metrics, we also used the phylogenetic endemism (PE) metric that calculates the fraction of branches restricted to specific regions. For this analysis, we used the phyloregion package in the R program via the function "results_PE<-phylo_endemism." PE identifies areas or communities that harbor restricted components of phylogenetic diversity and is a metric to aid conservation studies by establishing criteria to prioritize regions to be conserved based on evolutionary importance (parts of the phylogeny with limited spatial distribution) (Rosauer et al., 2009). The phylogenetic endemism (PE) index considers all the communities as the maximum spatial range. For example, if all species occur in all communities, the PE value would be 1, indicating low phylogenetic endemism.

2.5 | Statistical analysis

Phylogenetic structure analyses were performed using the 'picante' package, version 1.8.2 (Kembel et al., 2010). We used the "lm ()" function to conduct simple linear regressions and tested whether the three standardized phylogenetic metrics (ses.PD, ses.MPD, and ses.MNTD) were significantly related to flood amplitude, maximum flood height, and maximum water table depth. All analyses were performed in the program R, version 4.1.1 (R Core Team, 2023).

3 | RESULTS

3.1 | Tree inventory

The study included a total of 2,432 individuals, comprising 218 species, 133 genera, and 44 families. The phylogeny of all species is presented in the Data S1. Of the 2,432 individuals identified, 825 were situated in *terra-firme* riparian forests, 869 in intermediate positions, and 738 in *igapó* riparian forests. Among the 44 families recorded, the five most species-rich families were Fabaceae, Lecythidaceae,

Chrysobalanaceae, Sapotaceae, and Myrtaceae. Noteworthy among the most species-rich genera were *Eschweilera*, *Inga*, *Protium*, *Pouteria*, *Eugenia*, and *Buchenavia* (a list of species inventoried in the Falsino River is presented in Data S1).

Among the 10 most abundant species were *Euterpe oleracea* Mart., *Macrolobium bifolium* (Aubl.) Pers., *Myrcia umbraticola* (Kunth) E. Lucas & C.E. Wilson, *Licania micrantha* Miq., *Pterocarpus santalinoides* L'Hér. ex DC., *Cynometra hostmanniana* Tul., *Hymenopus macrophyllus* (Benth.) Sothers & Prance, *Discocarpus essequeboensis* Klotzsch, *Lecythis idatimon* Aubl., and *Leptobalanus sprucei* (Hook.f.) Sothers & Prance. Of the 218 species identified, 75 were represented by only one individual, while 29 species were represented by two individuals.

When examining family distribution along the hydrographic gradient, we observed that *terra-firme* forests presented a higher number of taxa with restricted spatial distribution, indicating the absence of close relatives in the community. The families Anacardiaceae, Arecaceae, Bignoniaceae, Goupiaceae, Lamiaceae, Lauraceae, Linaceae, Nyctaginaceae, Olacaceae, Stemonuraceae, and Urticaceae were exclusively present in these forests. Similarly, the families Polygonaceae and Putranjivaceae occurred exclusively in flooded forests.

3.2 | Hydrological gradient

In *terra-firme* small-streams riparian forests, the average flooding height was 116.33 cm, with a standard deviation (SD) of 44.11 cm. The mean of the maximum depth of the water table was -73 cm (± 40.42 cm). The average amplitude was approximately 189.33 cm (± 49.17 cm). In the riparian forests from the intermediate zone, the average flooding height was 456 cm (± 29.98), while the mean of the maximum depth of the water table was -138.67 cm (± 100.16 cm). The average amplitude was 594.67 cm (± 133.66 cm). In the *igapó* riparian forests, the average flooding height was 532.33 cm (± 46.39 cm), while the mean of the maximum depth of the water table was -183.5 cm (± 22.56 cm). The average amplitude was 715.17 cm (± 64.76 cm; Figure 3 and Table 1).

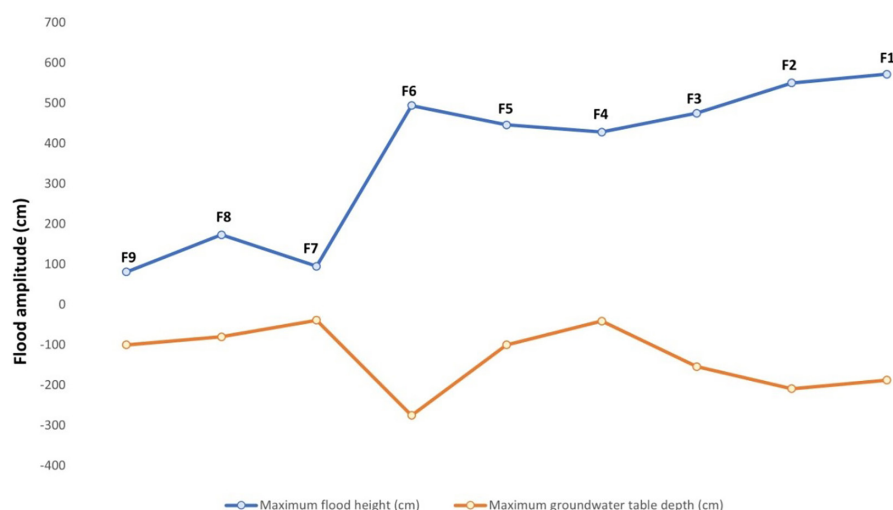


FIGURE 3 Mean of the annual variation of the hydrological gradient in the plots sampled in different stretches in the Amapá National Forest.

TABLE 1 Standardized effect size of phylogenetic diversity metrics (ses.PD), mean phylogenetic distance (ses.MPD), and the mean phylogenetic distance between the closest taxa (ses.MNTD), phylogenetic endemism (PE) index, and environmental variables observed in the plots of the Falsino River.

Plot	ses.PD	<i>p</i>	ses.MPD	<i>p</i>	ses.MNTD	<i>p</i>	PE	Alt	Mwt	Ampl
<i>terra-firme</i> riparian forests	1.58	.95	1.96	.84	1.83	.96	19.09	81	-100	181
<i>terra-firme</i> riparian forests	-1.02	.13	-0.38	.22	-0.79	.28	9.11	173	-80	253
<i>terra-firme</i> riparian forests	-1.28	.10	-1.29	.14	-1.05	.15	4.58	95	-39	134
Intermediate riparian forests	0.97	.23	-0.03	.24	-0.73	.25	5.21	494	-275	769
Intermediate riparian forests	-0.95	.20	-0.68	.44	-0.16	.47	4.09	446	-100	546
Intermediate riparian forests	-0.91	.25	-0.20	.44	-0.17	.46	4.25	428	-41	469
Foodplain riparian forests	0.60	.56	-0.73	.79	0.79	.80	5.89	475	-154	629
Foodplain riparian forests	0.99	.84	0.58	.74	0.68	.76	6.37	550	-209	759
Foodplain riparian forests	0.97	.32	-1.26	.45	-0.11	.48	5.68	572	-187.5	759.5

Note: Positive values of ses.PD, ses.MPD, and ses.MNTD suggest greater diversity and greater phylogenetic distance (superdispersion), while negative values suggest lower diversity and shorter phylogenetic distance (phylogenetic clustering). Values in bold indicate a significant variation. Abbreviations: Alt, flood height; Ampl, maximum flood amplitude; Plf, maximum depth of the water table.

3.3 | Species relationships and phylogenetic community structure

The phylogenetic structure analyses revealed that phylogenetic diversity varies along the flooding gradient. A portion of the *terra-firme* riparian forests exhibited positive ses.PD values, indicating greater phylogenetic diversity and phylogenetic overdispersion within the community. In this portion of the gradient, the ses.MNTD index demonstrated an excess of more recent species (Table 1). In contrast, in all other hydrographic portions, ses.MNTD values were close to zero, suggesting a random phylogenetic structure within each community. The *terra-firme* riparian zone presented the highest rate of phylogenetic endemism (PE), indicating a greater number of taxa with restricted spatial distribution (Table 1).

There is a significant reduction in phylogenetic diversity (ses.PD) with increasing water table depth ($p = .03$; $R^2 = 0.52$; Figure 4). There is no significant relationship between the flood amplitude and ses.PD ($p = .22$; $R^2 = 0.21$), ses.MPD ($p = .67$; $R^2 = 0.03$), and ses.MNTD ($p = .65$; $R^2 = 0.03$) indices. There is also no significant variation between the standardized values of richness and the maximum depth of the water table ($p = .69$; $R^2 = 0.02$), or the maximum height of aboveground flooding ($p = .15$; $R^2 = 0.02$) and flood amplitude ($p = .22$; $R^2 = 0.20$). No significant relationships were observed between maximum flood height, ses.PD ($p = .43$; $R^2 = 0.09$), ses.MPD ($p = .52$; $R^2 = 0.06$), and ses.MNTD ($p = .84$; $R^2 = 0.006$).

4 | DISCUSSION

In this study, we demonstrated the significant role of water table depth as a historical determinant in shaping the organization and structure of tree communities across the Amazonian landscape, spanning from *terra-firme* small streams to *igapó* floodplain riparian forests. By assessing the increasing influence of the flooding pulse,

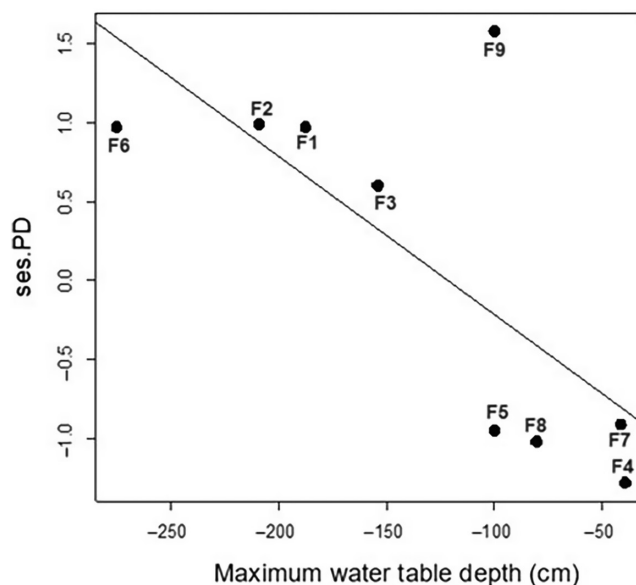


FIGURE 4 Linear regression between the phylogenetic diversity metrics (ses.PD) and water table depth.

we identified divergent patterns of phylogenetic diversity under varying hydrological conditions (Figure 4). In *terra-firme* uplands, characterized by shallower water tables and floods mainly caused by local rainfall, riparian forests exhibited higher species richness and phylogenetic diversity (ses.PD, Table 1 and Figure 4), leading to a phylogenetic overdispersion pattern, characterized by more recent relationships between species (Table 1). In contrast, other areas along the hydrographic gradient displayed a random distribution pattern and lower phylogenetic diversity and richness as water table depth increased, thus, supporting the hypothesis, proposed by Wittmann et al. (2010), which suggests that tree lineages were historically filtered by the flooding gradient to occupy similar niches of floodable areas.

The variation in species richness along the flooding gradient reflects complex interactions between flood levels and different plant survival strategies. The establishment of plants in Amazonian wetlands with water deficit or flooding depends on species tolerance and adaptation limits (Maurenza et al., 2009; Parolin, 2001). The phylogenetic approach employed here enabled us to identify ancient dominant lineages in the Amazonian lowlands, including Fabaceae, Lecythidaceae, Bignoniaceae, Sapotaceae, Chrysobalanaceae, and Burseraceae (Hoorn et al., 2010; Lewis, 2005; Lohmann et al., 2013; Luize et al., 2024). The dominant presence of Fabaceae and Lecythidaceae throughout the analyzed gradient underscores their capacities to adapt to diverse Amazonian environments and their fundamental role in sustaining ecosystems (Lavin et al., 2005). Other studies indicate the occurrence of families such as Chrysobalanaceae, Sapotaceae, and Bignoniaceae across a wide diversity of habitats, including different soil types like white sand, and experiencing various flood patterns (Bardon et al., 2013; Gentry, 1982; Jaramillo et al., 2010; Lamarre et al., 2012; Lohmann et al., 2013).

Comparing phylogenetic diversity (ses.PD) and phylogenetic endemism (PE) indices, we found a higher number of restricted lineages in *terra-firme* riparian forests, which also exhibited greater phylogenetic diversity (Table 1). High PE values were attributed to the presence of restricted families such as Anacardiaceae, Goupiaceae, Nyctaginaceae, and Stemonuraceae, each represented by a single species in this forest. This suggests that taxa adapted to these environment conditions may face challenges in dispersing to other habitats, emphasizing the importance of conserving forests harboring high concentrations of endemic taxa to preserve unique biological diversity and specific ecological niches, ensuring ecological integrity and ecosystem resilience. These results align with those of Rosauer et al. (2009), who observed the highest PE index in an area where a single species with no close relatives was found.

Our results further underscore the role of the water table in regulating tree communities by directly selecting species adapted to soil saturation (Cosme et al., 2017; Costa et al., 2023; Schietti et al., 2014). Accordingly, areas with deeper water tables tend to retain more water, limiting plant establishment and survival for species lacking adaptations to these environmental constraints (Stefanello et al., 2006). This limitation is evident in the reduction in phylogenetic diversity observed in *igapó* riparian forests. Conversely, areas with shallower water table tend to retain less water, reducing or eliminating establishment and survival constraints (Parolin & Wittmann, 2010; Piedade et al., 2013), resulting in increased community phylogenetic diversity, as observed in upland *terra-firme* riparian forests in this study. This decrease in species richness under more stressful hydro-edaphic conditions, such as increased flood duration, is consistent with those mentioned in several previous studies (i.e., Assis et al., 2015; Kubitzki, 1989; Moulatlet et al., 2014; Parolin, 2009; Piedade et al., 2013; Prance, 1979; Scudeller & de Souza, 2009). Furthermore, the observed phylogenetic patterns indicate adaptive convergence of different lineages for diversified use of ecological space, influencing community composition and species coexistence (Valencia et al., 1994; de Oliveira and Mori, 1999).

Studies demonstrated that floodable forests in the Amazon basin favor species adapted to frequent disturbances (Wittmann et al., 2004; Wittmann & Parolin, 2005), leading to species filtering along the flooding gradient.

The Amazon is considered one of the most important centers of plant diversification in the world (Antonelli et al., 2018). Although some species may have a wide distribution in the Amazon basin, most of them are locally restricted, supporting the environmental complexity of the region (Pitman et al., 2001; ter Steege et al., 2006, 2013). In this study, phylogenetic endemism revealed higher values in *terra-firme* riparian forests, which also exhibited the greatest phylogenetic diversity (ses.PD). High ses.PD values found in upland riparian forests align with the premise that phylogenetic diversity is greater in communities composed of evolutionarily distinct lineages (Cianciaruso et al., 2009; Magurran, 2004).

These results are consistent with those of studies describing how wetland plant community patterns are mainly influenced by environmental gradients and evolutionary relationships between different tree lineages (Dexter et al., 2017; Fine & Kembel, 2011; Keddy, 1992; Kraft & Ackerly, 2010). The prevalence of several tree families across the Amazon basin during evolutionary time scales reflects, even on a local scale, that independent and distinctive lineages adapted to distinct hydrological gradients, as previously reported for edaphic and topographic gradients (Nelson, 1992; Quesada et al., 2004). Our results support Gentry's (1982) hypothesis regarding habitat specialization driving the diversification of plant groups in the Amazon region. These are also consistent with multidisciplinary studies that state that the current flora of the Amazon is the result of complex ecological, geological, climatic, and hydrological events that acted on landscapes over millions of years (Cracraft et al., 2020; Gopal, 2009; Haffer, 1969; Haffer & Prance, 2002; Hoorn et al., 2010; Kubitzki, 1987; Pennington et al., 2004; Prance, 1985).

Terra-firme forests cover approximately 60% of the Amazon region and harbor significant knowledge gaps regarding its biodiversity (Carvalho et al., 2023; de Oliveira and Mori, 1999; Tuomisto et al., 2003). In this study, the highest phylogenetic diversity and occurrence of families with restricted spatial distribution were observed in riparian forests located along *terra-firme* low-order rivers of the drainage system, indicating greater ecosystem stability for this physiognomy (Cadotte et al., 2011). This underscores the urgency of conserving habitats conducive to the colonization of specific ecological niches. Simultaneously, the limited representation of families in regularly flooded forest environments, which comprise about 10% of the Amazon region, highlights the relevance of preserving these unique habitats. Despite limited studies on diversity and phylogenetic structure of plant communities in flooded forests, these areas face significant anthropogenic and climatic pressures (Schöngart et al., 2021).

Considering the anthropogenic climate change's impact on rainfall regimes and flood patterns of Amazon rivers, we anticipate that drastic changes in hydrology may lead to local extinction of habitat specialist tree species and consequently to a significant loss of phylogenetic diversity. Variations in precipitation patterns influence

flooding, affecting plant ability to disperse seeds in water. Species adapted to different flooding conditions may gain competitive advantages (Da Silva et al., 2023), thereby shaping the phylogenetic composition of communities over time. River continuum concept (Vannote et al., 1980) emphasizes considering the ecological integrity of the entire rivers and their watersheds for effective management. The spatial arrangement of tree phylogenetic diversity and endemism at local scales emphasizes the need to integrate watersheds into conservation frameworks to protect high phylogenetic diversity and endemism. This study's findings reinforce the critical importance of conserving and researching terrestrial and wetland ecosystems in the Amazon basin, as these areas are home to an invaluable wealth of biodiversity and play a fundamental role in maintaining biodiversity at local, regional, and global scales.

AUTHOR CONTRIBUTIONS

Sthefanie do Nascimento Gomes de Souza: Data curation (equal); formal analysis (equal); investigation (lead); writing – original draft (lead); writing – review and editing (lead). **Darlisson Mesquita Batista:** Data curation (equal); formal analysis (equal); writing – review and editing (equal). **Adriano Costa Quaresma:** Data curation (equal); writing – original draft (equal); writing – review and editing (equal). **Ana Luíza Costa:** Data curation (equal); writing – review and editing (equal). **Layon Oreste Demarchi:** Methodology (supporting); writing – review and editing (equal). **Bianca Weiss Albuquerque:** Methodology (supporting); writing – review and editing (equal). **Viviane Pagnussat Klein:** Methodology (supporting); writing – review and editing (equal). **Gildo Feitoza:** Data curation (equal); writing – review and editing (equal). **Angélica Faria de Resende:** Methodology (supporting); writing – review and editing (equal). **Gisele Biem Mori:** Methodology (supporting); writing – review and editing (equal). **Florian Wittmann:** Writing – original draft (supporting). **Leidiane Leão Oliveira:** Data curation (equal); writing – review and editing (equal). **Amanda Frederico Mortati:** Conceptualization (lead); data curation (equal); writing – review and editing (equal). **Alan Cavalcanti da Cunha:** Data curation (equal); funding acquisition (lead); writing – review and editing (equal). **Jochen Schongart:** Funding acquisition (lead); writing – review and editing (equal). **Aline Lopes:** Writing – original draft (equal); writing – review and editing (equal). **Maria Teresa Fernandez Piedade:** Conceptualization (equal); funding acquisition (lead); supervision (lead); writing – original draft (equal); writing – review and editing (equal). **Thiago André:** Conceptualization (equal); funding acquisition (lead); supervision (equal); writing – original draft (equal); writing – review and editing (equal).

ACKNOWLEDGMENTS

We thank the Coordination for the Improvement of Higher Education Personnel (CAPES) for granting the master's scholarship to SNGS. The University of Brasília (UnB) for financing publication fees (DPI/DPG/BCE 01/2024). This study is part of the Rede Ripária Project "Patterns and Structuring Processes of Biodiversity in Amazonian Wetlands," a partnership between the National Institute for Amazon Research (INPA), the Federal University of Western Pará (UFOPA),

and the Federal University of Amapá (UNIFAP) and the cooperation between the MAUA Group (Ecology, monitoring and sustainable use of wetlands) of INPA and the Max-Planck Institute of Biogeochemistry (INPA/Max-Planck Project). We acknowledge the funding by the National Council for Scientific and Technological Development (CNPq) (MCTI/CNPq Call No. 23/2017 Support for Biodiversity Research Networks in the Legal Amazon) of the Rede Ripária project (grant number 441462/2017-0) and its subprojects (grant numbers 441498/2017-5, 441481/2017-5, and 441520/2017-0). AFR thanks the São Paulo Research Foundation – FAPESP financial support (grant numbers: 22/14605-0 and 19/24049-5). AL thanks the Cesumar Institute of Science, Technology, and Innovation (ICETI) for the research grant.


CONFLICT OF INTEREST STATEMENT

The authors declare that there is no conflict of interest.

DATA AVAILABILITY STATEMENT

All relevant data are mentioned in the document and its supporting information files.

ORCID

Sthefanie do Nascimento Gomes de Souza  <https://orcid.org/0000-0002-5608-8852>

REFERENCES

- Antonelli, A., & Sanmartín, I. (2011). Why are there so many plant species in the Neotropics? *Taxon*, 60, 403–414.
- Antonelli, A., Zizka, A., Carvalho, F. A., Scharn, R., Bacon, C. D., Silvestro, D., & Condamine, F. L. (2018). Amazonia is the primary source of Neotropical biodiversity. *Proceedings of the National Academy of Sciences*, 115(23), 6034–6039.
- Assis, R. L., Haugaasen, T., Schöngart, J., Montero, J. C., Piedade, M. T., & Wittmann, F. (2015). Patterns of tree diversity and composition in Amazonian floodplain paleo-várzea forest. *Journal of Vegetation Science*, 26(2), 312–322.
- Ayres, J. M. C. (1986). *Uakaris and Amazonian flooded forest* (p. 380p). University of Cambridge.
- Bardon, L., Chamagne, J., Dexter, K. G., Sothers, C. A., Prance, G. T., & Chave, J. (2013). Origin and evolution of Chrysobalanaceae: Insights into the evolution of plants in the Neotropics: Phylogenetics of Chrysobalanaceae. *Botanical Journal of the Linnean Society*, 171, 19–37.
- Belúcio, L. P. (2020). Pulso hidrológico e gradientes edáficos em paisagens ripárias amazônicas. Dissertação (Mestrado em Ciências Ambientais) – Universidade Federal do Amapá, Fundação Nacional da Saúde.
- Bredin, Y. K., Hawes, J. E., Peres, C. A., & Haugaasen, T. (2020). Structure and composition of Terra Firme and seasonally flooded Várzea forests in the Western Brazilian Amazon. *Forests*, 11, 1361.
- Cadotte, M. W., Cardinale, B. J., & Oakley, T. H. (2008). Evolutionary history and the effect of biodiversity on plant productivity. *Proceedings National Academy of Sciences. United States of America*, 105, 17012–17017.
- Cadotte, M. W., Carscadden, K., & Mirotchnick, N. (2011). Beyond species: Functional diversity and the maintenance of ecological processes and services. *Journal of Applied Ecology*, 48, 1079–1087.
- Cadotte, M. W., Jonathan Davies, T., Regetz, J., Kembel, S. W., Cleland, E., & Oakley, T. H. (2010). Phylogenetic diversity metrics for ecological

- communities: Integrating species richness, abundance and evolutionary history. *Ecology Letters*, 13, 96–105.
- Carvalho, R. L., Resende, A. F., Barlow, J., França, F. M., Moura, M. R., Maciel, R., Alves-Martins, F., Shutt, J., Nunes, C. A., Elias, F., Silveira, J. M., Stegmann, L., Baccaro, F. B., Juen, L., Schietti, J., Aragão, L., Berenguer, E., Castello, L., Costa, F. R. C., ... Feitosa, Y. (2023). Pervasive gaps in Amazonian ecological research. *Current Biology*, 33, 3495–3504.e4.
- Cienciaruso, M. V., Batalha, M. A., Gaston, K. J., & Petchey, O. L. (2009). Including intraspecific variability in functional diversity. *Ecology*, 90, 81–89.
- Cosme, L. H. M., Schietti, J., Costa, F. R. C., & Oliveira, R. S. (2017). The importance of hydraulic architecture to the distribution patterns of trees in a central Amazonian forest. *New Phytologist*, 215, 113–125.
- Costa, F. R. C., Schietti, J., Stark, S. C., & Smith, M. N. (2023). The other side of tropical forest drought: Do shallow water table regions of Amazonia act as large-scale hydrological refugia from drought? *New Phytologist*, 237, 714–733.
- Cracraft, J., Ribas, C. C., d'Horta, F. M., Bates, J., Almeida, R. P., Aleixo, A., Boubli, J. P., Campbell, K. E., Cruz, F. W., Ferreira, M., & Fritz, S. C. (2020). The origin and evolution of Amazonian species diversity. In V. Rull & A. C. Carnaval (Eds.), *Neotropical diversification: Patterns and processes* (pp. 225–244). Springer International Publishing.
- Da Silva, E. B., Crizanto, J. L. P., de Abreu, C. H. M., dos Santos, E. S., Cunha, H. F. A., Brito, A. U., & da Cunha, A. C. (2023). Experimentation, modelling, and simulation of hydrochory in an Amazonian river. *Freshwater Biology*, 68(1), 141–154.
- de Oliveira, A. A., & Mori, S. A. (1999). A central Amazonian terra firme forest. I. High tree species richness on poor soils. *Biodiversity and Conservation*, 8, 1219–1244.
- Dexter, K. G., Lavin, M., Torke, B. M., Twyford, A. D., Kursar, T. A., Coley, P. D., Drake, C., Hollands, R., & Pennington, R. T. (2017). Dispersal assembly of rain forest tree communities across the Amazon basin. *Proceedings of the National Academy of Sciences*, 114, 2645–2650.
- Dirzo, R., & Raven, P. H. (2003). Global state of biodiversity and loss. *Annual Review of Environment and Resources*, 28, 137–167.
- dos Santos, E. S., Pinheiro Lopes, P. P., da Silva Pereira, H. H., de Oliveira Nascimento, O., Rennie, C. D., da Silveira, L., O'Reilly Sternberg, L., & da Cunha, A. C. (2018). The impact of channel capture on estuarine hydro-morphodynamics and water quality in the Amazon delta. *Science of the Total Environment*, 624, 887–899.
- Drucker, D. P., Costa, F. R. C., & Magnusson, W. E. (2008). How wide is the riparian zone of small streams in tropical forests? A test with terrestrial herbs. *Journal of Tropical Ecology*, 24, 65–74.
- Faith, D. P. (1992). Conservation evaluation and phylogenetic diversity. *Biological Conservation*, 61, 1–10.
- Faith, D. P., Reid, C. A. M., & Hunter, J. (2004). Integrating phylogenetic diversity, complementarity, and endemism for conservation assessment. *Conservation Biology*, 18, 255–261.
- Ferreira, C. D. S., Piedade, M. T. F., & Bonates, L. C. (2006). Germinação de sementes e sobrevivência de plântulas de *Himatanthus sucuuba* (Spruce) Wood. em resposta ao alagamento, nas várzeas da Amazônia Central. *Acta Amazonica*, 36, 413–418.
- Ferreira, L. V., Almeida, S. S., Amaral, D. D., & Parolin, P. (2005). Riqueza e composição de espécies da floresta de igapó e várzea da Estação Científica Ferreira Penna: Subsídios para o plano de Manejo da Floresta Nacional de Caxiuanã. *Pesquisas, Botânica*, 56, 103–116.
- Ferreira, L. V., de Almeida, S. S., & Parolin, P. (2010). Amazonian white- and blackwater floodplain forests in Brazil: Large differences on a small scale. *Ecotropica*, 16, 31–41.
- Ferrier, S., Pressey, R. L., & Barrett, T. W. (2000). A new predictor of the irreplaceability of areas for achieving a conservation goal, its application to real-world planning, and a research agenda for further refinement. *Biological Conservation*, 93, 303–325.
- Fine, P. V. A., & Kembel, S. W. (2011). Phylogenetic community structure and phylogenetic turnover across space and edaphic gradients in western Amazonian tree communities. *Ecography*, 34, 552–565.
- Fittkau, E. J. (1971). Distribution and ecology of amazonian chironomids (Diptera). *Canadian Entomologist*, 103, 407–413.
- Forest, F., Grenyer, R., Rouget, M., Davies, T. J., Cowling, R. M., Faith, D. P., Balmford, A., Manning, J. C., Procheş, Ş., van der Bank, M., Reeves, G., Hedderson, T. A. J., & Savolainen, V. (2007). Preserving the evolutionary potential of floras in biodiversity hotspots. *Nature*, 445, 757–760.
- Fukami, T., Martijn Bezemer, T., Mortimer, S. R., & Van Der Putten, W. H. (2005). Species divergence and trait convergence in experimental plant community assembly. *Ecology Letters*, 8, 1283–1290.
- Gentry, A. H. (1982). Patterns of Neotropical plant species diversity. In M. K. Hecht, B. Wallace, & G. T. Prance (Eds.), *Evolutionary biology* (pp. 1–84). Springer.
- Gerhold, P., Carlucci, M. B., Procheş, Ş., & Prinzing, A. (2018). The deep past controls the phylogenetic structure of present, local communities. *Annual Review of Ecology, Evolution, and Systematics*, 49, 477–497.
- Gómez, J., Schobbenhaus, C., & Montes, N. E. (2019). Geological map of South America at a scale of 1: 5M. Commission for the Geological map of the world (CGMW), Colombian geological survey, and geological survey of Brazil.
- Gopal, B. (2009). Biodiversity in wetlands. In E. Maltby & T. Barker (Eds.), *The wetlands handbook* (1st ed., pp. 65–95). Wiley.
- Gotelli, N. J., & Graves, G. R. (1996). *Null models in ecology*. Smithsonian Institution Press.
- Gregory, S. V., Swanson, F. J., McKee, W. A., & Cummins, K. W. (1991). An ecosystem perspective of riparian zones. *Bioscience*, 41, 540–551.
- Haffer, J. (1969). Speciation in Amazonian Forest birds: Most species probably originated in forest refuges during dry climatic periods. *Science*, 165, 131–137.
- Haffer, J., & Prance, G. T. (2002). Impulsos climáticos da evolução na Amazônia durante o Cenozóico: Sobre a teoria dos Refúgios da diferenciação biótica. *Estudos Avançados*, 16, 175–206.
- Hardy, O. J. (2008). Testing the spatial phylogenetic structure of local communities: Statistical performances of different null models and test statistics on a locally neutral community. *Journal of Ecology*, 96(5), 914–926.
- Hoorn, C., Bogotá-A, G. R., Romero-Baez, M., Lammertsma, E. I., Flantua, S. G. A., Dantas, E. L., Dino, R., do Carmo, D. A., & Chemale, F., Jr. (2017). The Amazon at sea: Onset and stages of the Amazon River from a marine record, with special reference to Neogene plant turnover in the drainage basin. *Global and Planetary Change*, 153, 51–65.
- Hoorn, C., Wesselingh, F. P., ter Steege, H., Bermudez, M. A., Mora, A., Sevink, J., Sanmartín, I., Sanchez-Meseguer, A., Anderson, C. L., Figueiredo, J. P., Jaramillo, C., Riff, D., Negri, F. R., Hooghiemstra, H., Lundberg, J., Stadler, T., Särkinen, T., & Antonelli, A. (2010). Amazonia through time: Andean uplift, climate change, landscape evolution, and biodiversity. *Science*, 330, 927–931.
- Householder, J. E., Wittmann, F., Schöngart, J., Piedade, M. T. F., Junk, W. J., Latrubesse, E. M., Quaresma, A. C., Demarchi, L. O., Aguiar, D. P., & Assis, R. L. (2024). One sixth of Amazonian tree diversity is dependent on river floodplains. *Nature Ecology & Evolution*, 8, 1–11.
- ICMBIO. (2014). *Plano de Manejo da Floresta Nacional do Amapá, Volume I - Diagnóstico*. Instituto Chico Mendes de Conservação da Biodiversidade.
- ICMBIO. (2016). *Plano de Manejo da Floresta Nacional do Amapá, Volume II - Planejamento*. Instituto Chico Mendes de Conservação da Biodiversidade.
- Irion, G., Junk, W. J., & de Mello, J. A. S. N. (1997). *The large central Amazonian river floodplains near Manaus: Geological, climatological, hydrological and geomorphological aspects* (pp. 23–46). Springer.
- Isbell, F., Balvanera, P., Mori, A. S., He, J., Bullock, J. M., Regmi, G. R., Seabloom, E. W., Ferrier, S., Sala, O. E., Guerrero-Ramírez, N. R., &

- Tavella, J. (2023). Expert perspectives on global biodiversity loss and its drivers and impacts on people. *Frontiers in Ecology and the Environment*, 21, 94–103.
- Jaramillo, C., Ochoa, D., Contreras, L., Pagani, M., Carvajal-Ortiz, H., Pratt, L. M., Krishnan, S., Cardona, A., Romero, M., Quiroz, L., Rodriguez, G., Rueda, M. J., de la Parra, F., Morón, S., Green, W., Bayona, G., Montes, C., Quintero, O., Ramirez, R., ... Vervoort, J. (2010). Effects of rapid global warming at the Paleocene-Eocene boundary on neotropical vegetation. *Science*, 330, 957–961.
- Jin, Y., & Qian, H. (2019). VPhyloMaker: An R package that can generate very large phylogenies for vascular plants. *Ecography*, 42, 1353–1359.
- Junk, W. J., Bayley, P. B., & Sparks, R. E. (1989). The flood pulse in river-floodplain systems. *Proceedings of the International Large River Symposium*, 106, 110–127.
- Junk, W. J., & Piedade, M. T. F. (1997). Plant life in the floodplain with special reference to herbaceous plants. In W. J. Junk (Ed.), *The Central Amazon floodplain: Ecology of a pulsing system* (pp. 147–185). Springer.
- Junk, W. J., Piedade, M. T. F., Schöngart, J., Cohn-Haft, M., Adeney, J. M., & Wittmann, F. (2011). A classification of major naturally-occurring Amazonian lowland wetlands. *Wetlands*, 31, 623–640.
- Junk, W. J., Wittmann, F., Schöngart, J., & Piedade, M. T. F. (2015). A classification of the major habitats of Amazonian black-water river floodplains and a comparison with their white-water counterparts. *Wetlands Ecology and Management*, 23, 677–693.
- Keddy, P. A. (1992). Assembly and response rules: Two goals for predictive community ecology. *Journal of Vegetation Science*, 3, 157–164.
- Kembel, S. W., Cowan, P. D., Helmus, M. R., Cornwell, W. K., Morlon, H., Ackerly, D. D., Blomberg, S. P., & Webb, C. O. (2010). Picante: R tools for integrating phylogenies and ecology. *Bioinformatics*, 26, 1463–1464.
- Koga, H., Ikematsu, S., & Kimura, S. (2024). Diving into the water: Amphibious plants as a model for investigating plant adaptations to aquatic environments. *Annual Review of Plant Biology*, 75, 24919.
- Kozłowski, T. T. (1997). Responses of woody plants to flooding and salinity. *Tree Physiology*, 17, 490.
- Kraft, N. J. B., & Ackerly, D. D. (2010). Functional trait and phylogenetic tests of community assembly across spatial scales in an Amazonian forest. *Ecological Monographs*, 80, 401–422.
- Kubitzki, K. (1987). Phenylpropanoid metabolism in relation to land plant origin and diversification. *Journal of Plant Physiology*, 131, 17–24.
- Kubitzki, K. (1989). The ecogeographical differentiation of Amazonian inundation forests. *Plant Systematics and Evolution*, 162, 285–304.
- Lamarre, G. P. A., Baraloto, C., Fortunel, C., Dávila, N., Mesones, I., Rios, J. G., Rios, M., Valderrama, E., Pilco, M. V., & Fine, P. V. A. (2012). Herbivory, growth rates, and habitat specialization in tropical tree lineages: Implications for Amazonian beta-diversity. *Ecology*, 93, S195–S210.
- Lavin, M., Herendeen, P. S., & Wojciechowski, M. F. (2005). Evolutionary rates analysis of Leguminosae implicates a rapid diversification of lineages during the tertiary. *Systematic Biology*, 54, 575–594.
- Lewis, G. (2005). *Legumes of the world*. Kew Publishing.
- Lobo, G., Wittmann, F., & Piedade, M. T. F. (2019). Response of black-water floodplain (igapó) forests to flood pulse regulation in a dammed Amazonian river. *Forest Ecology and Management*, 434, 110–118.
- Lobo, P. C., & Joly, C. A. (1995). Mecanismos de tolerância à inundação de plantas de Talauma ovata St. Hil. (Magnoliaceae), uma espécie típica de matas de brejo. *Revista Brasileira de Botânica*, 18, 177–183.
- Lohmann, L. G., Bell, C. D., Calió, M. F., & Winkworth, R. C. (2013). Pattern and timing of biogeographical history in the Neotropical tribe Bignonieae (Bignoniaceae): Bignonieae biogeography. *Botanical Journal of the Linnean Society*, 171, 154–170.
- Luize, B. G., Bauman, D., Ter Steege, H., Palma-Silva, C., Do Amaral, I. L., De Souza Coelho, L., de Almeida Matos, F. D., de Andrade Lima Filho, D., Salomão, R. P., Wittmann, F., & Castilho, C. V. (2024). Geography and ecology shape the phylogenetic composition of Amazonian tree communities. *Journal of Biogeography*, 1–22.
- Magurran, A. E. (2004). *Measuring biological diversity* (p. 272). Blackwell Publishers.
- Maurenza, D., Marenco, R. A., & Piedade, M. T. F. (2009). Efeito da inundação de longa duração sob o crescimento de Pouteria glomerata (Sapotaceae), uma arbórea da várzea da Amazônia Central. *Acta Amaz*, 39, 519–526.
- Miller, E. T., Farine, D. R., & Trisos, C. H. (2017). Phylogenetic community structure metrics and null models: A review with new methods and software. *Ecography*, 40, 461–477.
- Moulatlet, G. M., Costa, F. R. C., Rennó, C. D., Emilio, T., & Schietti, J. (2014). Local hydrological conditions explain floristic composition in lowland Amazonian forests. *Biotropica*, 46, 395–403.
- Muscarella, R., Emilio, T., Phillips, O. L., Lewis, S. L., Slik, F., Baker, W. J., Couvreur, T. L. P., Eiserhardt, W. L., Svenning, J. C., Affum-Baffoe, K., Aiba, S. I., de Almeida, E. C., de Almeida, S. S., de Oliveira, E. A., Álvarez-Dávila, E., Alves, L. F., Alvez-Valles, C. M., Carvalho, F. A., Guarín, F. A., ... Balslev, H. (2020). The global abundance of tree palms. *Global Ecology and Biogeography*, 29, 1495–1514.
- Myers, N., Mittermeier, R. A., Mittermeier, C. G., Da Fonseca, G. A. B., & Kent, J. (2000). Biodiversity hotspots for conservation priorities. *Nature*, 403, 853–858.
- Naiman, R. J., & Décamps, H. (1997). The ecology of interfaces: Riparian zones. *Annual Review of Ecology and Systematics*, 28, 621–658.
- Nelson, B. W. (1992). Diversidade florística de ecossistemas amazônicos. *Revista do Instituto Florestal*, 4, 111–118.
- Owen, N. R., Gumbs, R., Gray, C. L., & Faith, D. P. (2019). Global conservation of phylogenetic diversity captures more than just functional diversity. *Nature Communications*, 10, 859.
- Parolin, P. (2001). Morphological and physiological adjustments to waterlogging and drought in seedlings of Amazonian floodplain trees. *Oecologia*, 128, 326–335.
- Parolin, P. (2009). Submerged in darkness: Adaptations to prolonged submergence by woody species of the Amazonian floodplains. *Annals of Botany*, 103, 359–376.
- Parolin, P., & Wittmann, F. (2010). Struggle in the flood: Tree responses to flooding stress in four tropical floodplain systems. *AoB Plants*, 2010, plq003.
- Pausas, J. G., & Verdú, M. (2010). The jungle of methods for evaluating phenotypic and phylogenetic structure of communities. *Bioscience*, 60, 614–625.
- Pennington, D. W., Potting, J., Finnveden, G., Lindeijer, E., Joliet, O., Rydberg, T., & Rebitzer, G. (2004). Life cycle assessment part 2: Current impact assessment practice. *Environment International*, 30, 721–739.
- Piedade, M. T. F., Schöngart, J., Wittmann, F., Parolin, P., & Junk, W. J. (2013). Impactos ecológicos da inundação e seca na vegetação das áreas alagáveis amazônicas. *Eventos climáticos extremos na Amazônia: causas e consequências*, 409–461.
- Pitman, N. C. A., Terborgh, J. W., Silman, M. R., Núñez, V., P., Neill, D. A., Cerón, C. E., Palacios, W. A., & Aulestia, M. (2001). Dominance and distribution of tree species in upper amazonian terra firme forests. *Ecology*, 82, 2101–2117.
- Prance, G. (1985). Amazonia. In G. T. Prance & T. E. Lovejoy (Eds.), *Key environments* (1st ed., p. 442). Pergamon Press.
- Prance, G. T. (1979). Notes on the vegetation of Amazonia III. The terminology of Amazonian Forest types subject to inundation. *Brittonia*, 31, 26.
- Prance, G. T. (1980). A terminologia dos tipos de florestas amazônicas sujeitas a inundação. *Acta Amazonica*, 10, 499–504.

- Pressey, R. L., Johnson, I. R., & Wilson, P. D. (1994). Shades of irreplaceability: Towards a measure of the contribution of sites to a reservation goal. *Biodiversity and Conservation*, 3, 242–262.
- Quesada, C. A., Miranda, A. C., Hodnett, M. G., Santos, A. J. B., Miranda, H. S., & Breyer, L. M. (2004). Seasonal and depth variation of soil moisture in a burned open savanna (Campo Sujo) in central Brazil. *Ecological Applications*, 14, 33–41.
- R Core Team. (2023). R: A language and environment for statistical computing. In R. Relyea & R. Ricklefs (Eds.), *A Economia Da Natureza* (8th ed.). Guanabara Koogan.
- Ricklefs, R. E., Relyea, R., & Richter, C. (2014). *Ecology: the economy of nature* (Vol. 7). WH Freeman.
- Ricotta, C. (2005). Through the jungle of biological diversity. *Acta Biotheoretica*, 53, 29.
- Rocha, M., Assis, R. L., Piedade, M. T. F., Feitosa, Y. O., Householder, J. E., Lobo, G. D. S., Demarchi, L. O., Albuquerque, B. W., Quaresma, A. C., Ramos, J. F., & Schöngart, J. (2019). Thirty years after Balbina dam: Diversity and floristic composition of the downstream floodplain forest, Central Amazon, Brazil. *Ecohydrology*, 12, e2144.
- Rosauer, D., Laffan, S. W., Crisp, M. D., Donnellan, S. C., & Cook, L. G. (2009). Phylogenetic endemism: A new approach for identifying geographical concentrations of evolutionary history. *Molecular Ecology*, 18, 4061–4072.
- Salati, E., & Vose, P. B. (1984). Amazon Basin: A system in equilibrium. *Science*, 225, 129–138.
- Sandel, B. (2018). Richness-dependence of phylogenetic diversity indices. *Ecography*, 41, 837–844.
- Schietti, J., Emilio, T., Rennó, C. D., Drucker, D. P., Costa, F. R. C., Nogueira, A., Baccaro, F. B., Figueiredo, F., Castilho, C. V., Kinupp, V., Guillaumet, J. L., Garcia, A. R. M., Lima, A. P., & Magnusson, W. E. (2014). Vertical distance from drainage drives floristic composition changes in an Amazonian rainforest. *Plant Ecology and Diversity*, 7, 241–253.
- Schöngart, J., Wittmann, F., Faria De Resende, A., Assahira, C., De Sousa Lobo, G., Rocha Duarte Neves, J., da Rocha, M., Biem Mori, G., Costa Quaresma, A., Oreste Demarchi, L., & Weiss Albuquerque, B. (2021). The shadow of the Balbina dam: A synthesis of over 35 years of downstream impacts on floodplain forests in Central Amazonia. *Aquatic Conservation: Marine and Freshwater Ecosystems*, 31, 1117–1135.
- Scudeller, V. V., & de Souza, A. M. G. (2009). Florística da mata de igapó na Amazônia Central. In E. N. Santos-Silva & V. V. Scudeller (Eds.), *Biotupé: Meio Físico, Diversidade Biológica e Sociocultural Do Baixo Rio Negro, Amazônia Central* (Vol. 2, pp. 97–108). UEA Edições.
- Sioli, H. (1985). Amazônia. In J. Becker (Ed.), *Fundamentos de ecologia da maior região de florestas tropicais* (p. 72). Vozes.
- Smith, S. A., & Brown, J. W. (2018). Constructing a broadly inclusive seed plant phylogeny. *American Journal of Botany*, 105, 302–314.
- Sombroek, W. (2000). Amazon landforms and soils in relation to biological diversity. *Acta Amazonica*, 30, 81–100.
- Stefanello, R., Garcia, D. C., de Menezes, N. L., Muniz, M. F. B., & Wrasse, C. F. (2006). Efeito da luz, temperatura e estresse hídrico no potencial fisiológico de sementes de funcho. *Revista Brasileira de Sementes*, 28, 135–141.
- Stropp, J., Sleen, P. V. D., Assunção, P. A., Silva, A. L. D., & Steege, H. T. (2011). Tree communities of white-sand and terra-firme forests of the upper Rio Negro. *Acta Amazonica*, 41, 521–544.
- SUDAM. (1984). *Atlas Climatológico da Amazônia Brasileira* (p. 125p). SUDAM/PHCA.
- ter Steege, H., Pitman, N. C. A., Phillips, O. L., Chave, J., Sabatier, D., Duque, A., Molino, J. F., Prévost, M. F., Spichiger, R., Castellanos, H., von Hildebrand, P., & Vásquez, R. (2006). Continental-scale patterns of canopy tree composition and function across Amazonia. *Nature*, 443, 444–447.
- ter Steege, H., Pitman, N. C. A., Sabatier, D., Baraloto, C., Salomão, R. P., Guevara, J. E., Phillips, O. L., Castilho, C. V., Magnusson, W. E., Molino, J. F., Monteagudo, A., Núñez Vargas, P., Montero, J. C., Feldpausch, T. R., Coronado, E. N. H., Killeen, T. J., Mostacedo, B., Vasquez, R., Assis, R. L., ... Silman, M. R. (2013). Hyperdominance in the Amazonian tree flora. *Science*, 342, 1243092.
- Tucker, C. M., & Cadotte, M. W. (2013). Fire variability, as well as frequency, can explain coexistence between seeder and resprouter life histories. *Journal of Applied Ecology*, 50, 594–602.
- Tuomisto, H., Ruokolainen, K., Aguilar, M., & Sarmiento, A. (2003). Floristic patterns along a 43-km long transect in an Amazonian rain forest. *Journal of Ecology*, 91, 743–756.
- Valencia, R., Balslev, H., & Miño, P. Y. (1994). High tree alpha-diversity in Amazonian Ecuador. *Biodiversity and Conservation*, 3, 21–28.
- Vane-Wright, R. I., Humphries, C. J., & Williams, P. H. (1991). What to protect?—Systematics and the agony of choice. *Biological Conservation*, 55, 235–254.
- Vannote, R. L., Minshall, G. W., Cummins, K. W., Sedell, J. R., & Cushing, C. E. (1980). The river continuum concept. *Canadian Journal of Fisheries and Aquatic Sciences*, 37(1), 130–137.
- Webb, C. O. (2000). Exploring the phylogenetic structure of ecological communities: An example for rain Forest trees. *The American Naturalist*, 156, 145–155.
- Webb, C. O., Ackerly, D. D., McPeck, M. A., & Donoghue, M. J. (2002). Phylogenies and community ecology. *Annual Review of Ecology and Systematics*, 33, 475–505.
- Wittmann, F., Householder, E., Piedade, M. T. F., De Assis, R. L., Schöngart, J., Parolin, P., & Junk, W. J. (2013). Habitat specificity, endemism and the neotropical distribution of Amazonian white-water floodplain trees. *Ecography*, 36, 690–707.
- Wittmann, F., Householder, J. E., Piedade, M. T. F., Schöngart, J., Demarchi, L. O., Quaresma, A. C., & Junk, W. J. (2022). A review of the ecological and biogeographic differences of Amazonian floodplain forests. *Water*, 14(21), 3360.
- Wittmann, F., Junk, W. J., & Piedade, M. T. F. (2004). The várzea forests in Amazonia: Flooding and the highly dynamic geomorphology interact with natural forest succession. *Forest Ecology and Management*, 196, 199–212.
- Wittmann, F., & Parolin, P. (1999). Phenology of six tree species from central amazonian várzea. *Ecotropica*, 5, 51–57.
- Wittmann, F., & Parolin, P. (2005). Aboveground roots in Amazonian floodplain Trees1. *Biotropica*, 37, 609–619.
- Wittmann, F., Schöngart, J., & Junk, W. J. (2010). Phytogeography, species diversity, community structure and dynamics of central Amazonian floodplain forests. In W. J. Junk, M. T. F. Piedade, F. Wittmann, J. Schöngart, & P. Parolin (Eds.), *Amazonian floodplain forests* (Vol. 210, pp. 61–102). Springer.

SUPPORTING INFORMATION

Additional supporting information can be found online in the Supporting Information section at the end of this article.

How to cite this article: de Souza, S. d. N. G., Batista, D. M., Quaresma, A. C., Costa, A. L., Demarchi, L. O., Albuquerque, B. W., Klein, V. P., Feitoza, G., de Resende, A. F., Mori, G. B., Wittmann, F., Oliveira, L. L., Mortati, A. F., da Cunha, A. C., Schongart, J., Lopes, A., Piedade, M. T. F., & André, T. (2024). Soil flooding filters evolutionary lineages of tree communities in Amazonian riparian forests. *Ecology and Evolution*, 14, e11635. <https://doi.org/10.1002/ece3.11635>